

# **Part-per-trillion LC-MS/MS determination of neonicotinoids in small volumes of songbird plasma**

Chunyan Hao<sup>1\*</sup>, Margaret L. Eng<sup>2</sup>, Fengrong Sun<sup>1</sup>, and Christy A. Morrissey<sup>3,4</sup>

<sup>1</sup>Laboratory Services Branch, Ontario Ministry of the Environment and Climate Change, 125  
Resources Road, Etobicoke, Ontario M9P 3V6, Canada

<sup>2</sup>Toxicology Centre, University of Saskatchewan, 44 Campus Dr, Saskatoon, Saskatchewan, S7N  
5B3, Canada

<sup>3</sup>Department of Biology, University of Saskatchewan, 112 Science Place, Saskatoon,  
Saskatchewan, S7N 5E2, Canada

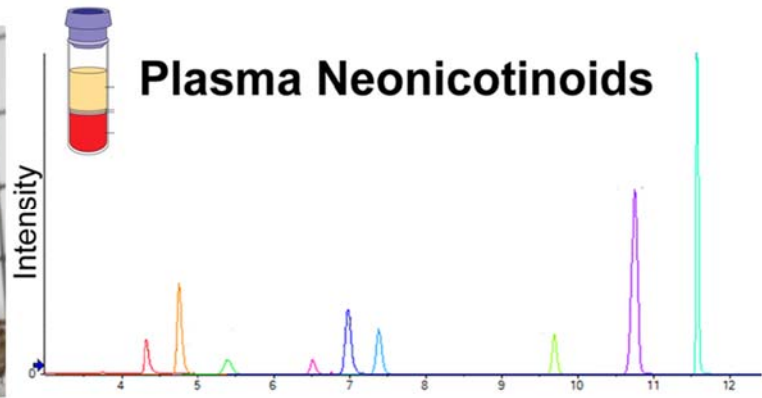
<sup>4</sup>School of Environment and Sustainability, University of Saskatchewan, 117 Science Place,  
Saskatoon, Saskatchewan, S7N 5C8, Canada

\*Correspondence Author.

Email: [chunyan.hao@ontario.ca](mailto:chunyan.hao@ontario.ca)

Tel: 416 235 6033

Fax: 416 235 5900



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### 3 **Highlights**

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- Non-lethal part-per-trillion LC/MS-MS method developed for the detection of neonicotinoids in songbird plasma

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- 50  $\mu$ L plasma samples extracted using simple precipitation and dilution procedure

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- First study to confirm widespread neonicotinoid exposure in free-living songbirds

8

- Method confirmed imidacloprid concentration increase in plasma after dosing

9 **Abstract:**

10           Neonicotinoids are the most widely used class of insecticides in the world, and there are  
11 increasing concerns about their effects on non-target organisms. Analytical methods to diagnose  
12 exposure to neonicotinoids in wildlife are still very limited, particularly for small animals such as  
13 songbirds. Blood can be used as a non-lethal sampling matrix, but the sample volume is limited  
14 by body size. Neonicotinoids have a low bioaccumulation potential and are rapidly metabolized,  
15 therefore, sensitive assays are critically needed to reliably detect their residues in blood samples.  
16 We developed an efficient LC-MS/MS method at a part-per-trillion (pg/mL) level to measure  
17 eight neonicotinoid related insecticides (acetamiprid, clothianidin, dinotefuran, flonicamid,  
18 imidacloprid, nitenpyram, thiacloprid and thiamethoxam) plus one metabolite (6-chloronicotinic  
19 acid) in small volumes (50 µl) of avian plasma. The average recovery of target compounds  
20 ranged from 95.7 to 101.3%, and relative standard deviations were between 0.82-2.13%. We  
21 applied the method to screen blood samples from 36 seed-eating songbirds (white-crowned  
22 sparrows; *Zonotrichia leucophrys*) at capture, and detected imidacloprid in 78% (28 of 36),  
23 thiamethoxam in 22% (8 of 36), thiacloprid in 11% (4 of 36), and acetamiprid in 11% (4 of 36)  
24 of wild-caught sparrows. Six hours after capture, birds were orally dosed with 0 (control), 1.2 or  
25 3.9 mg of imidacloprid/kg bw, test results using this method indicated that plasma imidacloprid  
26 was significantly elevated (low 26-times, high 316-times) in exposed groups. This is the first  
27 study to confirm neonicotinoid exposure in small free-living songbirds through non-lethal blood  
28 sampling, and to demonstrate that environmentally realistic doses significantly elevate  
29 circulating imidacloprid concentrations. This sensitive method could be applied to characterize  
30 exposure to neonicotinoids in free-living wildlife and in toxicological studies.

31 **Keywords:** Neonicotinoids, 6-Chloronicotinic Acid (6-CAN), Liquid Chromatography-Tandem  
32 Mass Spectrometry (LC-MS/MS), Multiple Reaction Monitoring (MRM), Blood  
33 Plasma  
34

## 35 **1. Introduction**

36

37       Neonicotinoids are a class of insecticides with nicotine-like molecular structure that have  
38 been used widely in the last two decades. They are registered for use in more than 120 countries  
39 and are the largest class of insecticides sold worldwide [1]. These insecticides are employed  
40 dominantly in insecticidal seed treatments. Neonicotinoids are considered less toxic to  
41 vertebrates and relatively safer for the environment [2, 3] due to their specific mode-of-action as  
42 insect-nicotinic acetylcholine receptor (nAChR) inhibitor [4]. However, there is increasing  
43 concern that the high availability of neonicotinoids in the environment could have negative direct  
44 or indirect effects on non-target invertebrate and vertebrate wildlife, including aquatic  
45 invertebrates [5, 6,7], bee colonies [8, 9, 10], insectivorous birds [11], and seed eating birds [12,  
46 13, 14, 15]. Several regulatory agencies have restricted the use of certain neonicotinoids, and  
47 Canada, the United States, and Europe are currently re-evaluating clothianidin, imidacloprid,  
48 thiamethoxam and their associated products [16, 17, 18].

49       Wildlife that use agricultural landscapes have the potential to be directly exposed to  
50 neonicotinoids through several exposure routes, including consumption of treated seeds. The  
51 timing of seeding for many crops in the northern mid-latitudes coincides with the spring  
52 migration in birds [19], and thus seed-eating birds that use cropland for refueling may be  
53 particularly susceptible. Evidence of direct exposure to neonicotinoids in seed-eating birds  
54 comes from post-mortem analysis of tissue residues and gastrointestinal contents [13, 20, 21] and  
55 from field observations [14]. There is a need for sensitive and effective analytical methods to  
56 non-lethally determine neonicotinoid concentrations in small birds. Blood is a common sample  
57 media to evaluate pesticide exposure [22], and can be safely used for sampling purposes, as long

58 as total volume extracted is limited to less than 10% blood volume (approx. 1% of body mass)  
59 per sample and less than 15% in a 14 day period for repeated sampling [23]. An increasing  
60 number of methods have been published for the determination of neonicotinoids in agricultural,  
61 food products and environmental samples [24, 25, 26, 27, 28, 29, 30, 31, 32, 33, 34] with liquid  
62 chromatography-tandem mass spectrometry (LC-MS/MS) as the dominant analytical  
63 methodology [35, 36, 37, 38, 39, 40, 41]. However, very little method development work has  
64 focused on non-lethal biotic matrices in wildlife. We only found one previous study that  
65 described an analytical procedure to determine neonicotinoids in bird blood samples with limits  
66 of quantification from 2000 to 10000 pg/mL, and requiring 500  $\mu$ L of blood [42]; however,  
67 considering the low bioaccumulation potential of neonicotinoids in mammals [43, 44], more  
68 sensitive methods are needed to reliably detect exposures, using smaller volumes suitable for  
69 animals such as songbirds.

70 We previously developed a direct aqueous injection LC-MS/MS method [45] for the  
71 analysis of eight neonicotinoids, including acetamiprid, imidacloprid, nitenpyram, thiacloprid  
72 (first generation neonicotinoids), clothianidin, thiamethoxam (second generation), dinotefuran  
73 (third generation) and flonicamid, a new systemic pesticide that is often included in the  
74 neonicotinoid group [46] in environmental water samples. A common metabolite of the first  
75 generation neonicotinoids, 6-chloronicotinic acid (6-CNA) [47, 48], was also included in the  
76 method as it had been detected in bees after exposure to neonicotinoids [35]. Compared to water  
77 samples, biological fluids like blood usually contain more complicated matrix components that  
78 may cause matrix effects during electrospray ionization [49]. Solid phase extraction with 96-well  
79 plates are often used to concentrate and cleanup this type of small volume biological samples  
80 [50]. QuEChERS (Quick, Easy, Cheap, Effective, Rugged, and Safe) technique is another

81 alternative where sugars, lipids, proteins, pigments, excess water, etc. can be removed by  
82 dispersive solid phase extraction (dSPE) after extraction [51]. The dilute-and-shoot approach  
83 coupled with LC-MS has also been applied to deal with blood plasma samples [52], and is the  
84 simplest approach as it uses direct aqueous injection; however, detection limits suffer from the at  
85 least 100-fold dilution required to reduce the matrix effect. Here, we combined quick  
86 deproteinization and small dilution ( $\leq 10$ -fold) to develop a simple method with good sensitivity  
87 to measure neonicotinoids and 6-CNA in avian plasma

88         The overall objectives of our current work were 1) to develop an efficient and sensitive  
89 LC-MS/MS approach to measure eight neonicotinoid related pesticides in bird plasma, 2) to  
90 apply the method to assess concentrations of neonicotinoids in wild-caught migrating seed-eating  
91 songbirds (Eastern white-crowned sparrow; *Zonotrichia leucophrys*) at capture, and 3) to verify  
92 expected increases in plasma concentrations in the same birds following controlled oral exposure  
93 to environmentally relevant levels of imidacloprid.

94

## 95 **2. Material and Methods**

96

### 97 **2.1. Chemicals & Reagents**

98         Neat standards of native and isotope-labelled neonicotinoids, American Chemical Society  
99 (ACS) reagent grade ammonium acetate ( $\text{CH}_3\text{COONH}_4$ ), formic acid ( $\text{HCOOH}$ ), distilled in  
100 glass grade methanol ( $\text{CH}_3\text{OH}$ ), and plasma from chicken and bovine (3.8% trisodium citrate as  
101 anticoagulant) were purchased from Sigma-Aldrich (Oakville, ON, Canada). High purity water  
102 was produced by passing reverse osmosis water through a Barnstead NANOpure<sup>TM</sup> water  
103 purification system (Mississauga, ON, Canada). The isotope-labelled internal standards (ILISs)

104 acetamiprid-d<sub>3</sub>, clothianidin-d<sub>3</sub>, imidacloprid-d<sub>4</sub> and thiamethoxam-d<sub>3</sub> were used to correct  
105 extraction loss, matrix effects and instrument variability and to monitor method performance in  
106 each sample. Working solutions for native target compounds and ILISs in methanol were  
107 prepared from neat standards as described previously [45]. A custom-mix stock solution of eight  
108 native neonicotinoids in acetonitrile was purchased from AccuStandard Inc (New Haven, CT,  
109 USA) for quality control. All standard solutions were stored at 5 ± 3°C and allowed to reach  
110 room temperature before use.

111

## 112 **2.2. Bird Blood Sample Collection, dosing, and captive housing**

113 White-crowned sparrows are seed-eating migratory songbirds that have the potential to  
114 be directly exposed to insecticides primarily through consumption of treated seeds or granules or  
115 by coming in contact with foliage and soils sprayed with neonicotinoids. In May 2017 as part of  
116 a companion study, a total of 36 Eastern white-crowned sparrows (*Z. l. leucophrys*) were  
117 captured in mist nets or sparrow traps at Long Point Bird Observatory, Ontario (42.5829° N,  
118 80.3985° W). Birds were blood sampled at capture, and held overnight in cages (66 cm L x 46  
119 cm W x 50 cm H) on site in an animal housing room within a mobile research laboratory, with 2  
120 to 3 birds per cage. Birds were weighed and dosed between 09:00 to 11:00, then weighed and  
121 blood sampled between 15:00 to 17:30. Mean (±SE) time between dosing and blood sampling  
122 was 6.1 ± 0.1 hours. Birds were randomly assigned to treatment groups, and dosing was through  
123 oral gavage directly into the crop using 20G curved stainless steel tube feeders at a volume of 10  
124 ml/kg bw, with either the low dose (1.2 mg IMI/kg bw; n = 12), high dose (3.9 mg IMI/kg bw; n  
125 = 12), or vehicle control (sunflower oil; n = 12). Avian plasma samples used as blanks for quality  
126 control were collected from Gambel's white-crowned sparrows (*Z. l. gambelii*) captured in



127 Saskatchewan in May 2016 for a previous study [12]. These birds had been held in captivity for  
128  $\geq 14$  days in the Facility for Applied Avian Research at the University of Saskatchewan. Blood-  
129 samples were taken from the brachial vein following puncture with a 26G needle. Blood was  
130 collected into heparinized tubes, centrifuged to separate plasma from red blood cells, and plasma  
131 was then stored frozen at  $-20^{\circ}\text{C}$  until analysis. Birds were provided with water and a mixture of  
132 millet, black oil sunflower seeds, and poultry starter crumbles (Proform 26%) *ad libitum*.  
133 Research protocols were in compliance with the Canadian Council on Animal Care guidelines  
134 and approved by the University of Saskatchewan Animal Care Committee (AUP 20110043), and  
135 conducted under Canadian Wildlife Service Scientific Permits 15SKSC005 and SC00008.

136

### 137 **2.3. Dosing solution composition and analysis**

138 Dosing solutions were made by dissolving technical grade imidacloprid (Sigma Aldrich  
139 37894) in a small volume of acetone, then diluting with food-grade organic sunflower oil  
140 (Compliments brand, Sobeys Canada). Nominal concentrations were 0 mg IMI/ml (control), 0.10  
141 mg IMI/ml (low), and 0.41 mg IMI/ml (high). Solutions were stirred overnight to evaporate off  
142 acetone, and stored in amber glass bottles in the dark for the duration of the study. Nominal  
143 concentrations were selected to be equivalent to 2.5% (low) and 10% (high) of the house sparrow  
144 LD50 (41 mg/kg bw) [53], where the high dose (4.0 mg/kg bw) was below that used in a  
145 previous captive study in white-crowned sparrows (10.25 mg/kg bw) [12]. These doses are  
146 environmentally realistic, with the mass consumed of imidacloprid in the low dose being  
147 equivalent to the mass of imidacloprid on  $\sim 1.1$  treated canola seeds or 1 wheat kernel, and the  
148 high dose is equivalent to the mass on  $\sim 3.5$  treated canola or 3.2 treated wheat kernels, according  
149 to current US application rates (Table S1).

150 Dosing solution concentrations of imidacloprid were confirmed by LC-MS/MS analyses  
151 at the National Hydrology Research Centre, Environment and Climate Change Canada,  
152 Saskatoon, SK. Solutions were diluted 100x into an intermediate solvent (acetone), then further  
153 diluted into water (20x for the control and low solutions, 100x for the high solution). Diluted  
154 aqueous samples were directly injected into the mass spectrometer (Waters 2695 Alliance HPLC  
155 system; Waters Corp., Milford, MA), using the same instrumentation and calibration methods  
156 described in Eng et al. 2017 [12]. Measured concentrations of imidacloprid were 0.12 mg/ml  
157 (low) and 0.39 mg/ml (high). Vehicle control oil and all blanks had no detectable levels of  
158 imidacloprid.

159

#### 160 **2.4. Sample Preparation and Instrument Analysis**

161 Plasma samples were warmed up to room temperature, 50 to 200  $\mu$ L of each sample was  
162 put into a 1.8 mL amber glass HPLC vial, then 20  $\mu$ L of ILIS spiking solution in methanol, 20  $\mu$ L  
163 of methanol and 2.5  $\mu$ L of formic acid were added. The vial contents were mixed well and the  
164 whole content was blown down to dryness with light nitrogen flow at 30  $^{\circ}$ C. Then 500  $\mu$ L of  
165 20:80 methanol:water was added into the vial to reconstitute, and the vial was centrifuged at  
166 4000rpm for 20 minutes. The acid, organic solvent and blow down precipitated out proteins and  
167 other larger biomolecules in the plasma, a process referred to as deproteinization. Each vial was  
168 carefully removed from the centrifuge and  $\sim$ 200  $\mu$ L of supernatant was transferred to a clean  
169 HPLC vial. Blank plasma samples were processed exactly the same way. For each fortified  
170 plasma sample, 20  $\mu$ L of native spiking solution was added instead of pure methanol.

171 Dilution factors (10, 5 and 2.5 with 500  $\mu$ L final volume) and temperature before  
172 blowdown (-20 or 45 $^{\circ}$ C) were optimized to reduce matrix effects using commercial bovine and

173 chicken plasma samples fortified with target compounds. The recoveries of fortified isotope-  
174 labelled neonicotinoids in 50, 100, 200  $\mu\text{L}$  of bovine plasma and 200  $\mu\text{L}$  of chicken plasma were  
175 assessed.

176 A dilution factor of 10 (50  $\mu\text{L}$  plasma in 500  $\mu\text{L}$  final volume) was used for validation of  
177 the method with fortified sparrow plasma, and for unknown sparrow samples. The optimized  
178 method was validated for songbird samples using pooled plasma from Gambel's white-crowned  
179 sparrows that had been held in captivity for  $\geq 14$  days. Plasma pools contained no detectable  
180 neonicotinoids. Target compound concentrations were calculated based on calibration curves  
181 established with 5-level calibration standards, which were plotted using a  $1/x$  weighting and  
182 linear regression with internal standard correction.

183 The instrument analysis was carried out on a Sciex (Concord, ON, Canada)  
184 QTRAP<sup>®</sup>5500 mass spectrometer coupled with a Shimadzu Prominence/20 series (Columbia,  
185 MD, USA) LC system. 90  $\mu\text{L}$  of instrument ready solutions were injected into a Phenomenex  
186 Kinetex Biphenyl 2.6 mm 100 x 4.6 mm LC column. Multiple reaction monitoring (MRM) data  
187 was acquired and processed using electrospray ionization mode with the Analyst 1.6.3 software  
188 and the Scheduled MRM<sup>™</sup> (sMRM) algorithm with a target cycle time of 800 ms and a  
189 detection window of 60 s for each transition. Instrument conditions and parameters used were  
190 described in detail previously [45]. The two MRM transitions and corresponding collision  
191 energies, the ILISs used for quantitation, and the calibration ranges for each target compound are  
192 listed in Table 1. A control standard of a different source from calibration standards was  
193 analyzed with each batch of samples. The accuracy of each compound in the control standard  
194 had to be within 70 to 130% range. The accuracy (recovery) of each spiked replicate was  
195 calculated by Analyst Software by comparing the calculated concentration with the spiked

196 concentration. The average recovery (Avg Rec) and the relative standard deviation (RSD) of the  
197 spiked replicates were calculated by Excel formulae. The method detection limits (MDLs) were  
198 calculated according to the U.S. Environmental Protection Agency protocol [54].

199

## 200 **2.5. Statistical analysis**

201 Statistical analysis was completed using SAS 9.4. Imidacloprid concentration was log  
202 transformed to meet assumptions of normality. For samples that were below the MDL, random  
203 values between zero and the MDL were substituted into the dataset for statistical analysis.  
204 Comparisons of imidacloprid concentrations between dose groups over time (pre-dosing at  
205 capture vs post-dosing) were made with a linear mixed model (proc MIXED), with bird ID as a  
206 repeated subject effect, and fixed effects of dose, time, and dose\*time. Tests for differences  
207 between means were adjusted for multiple comparisons using the Tukey-Kramer method.  
208 Significance level was set at  $\alpha = 0.05$ .

209

## 210 **3. Results and Discussion**

211

### 212 **3.1 Analytical Method Optimization**

213 We developed a sensitive part-per-trillion LC-MS/MS method to analyze small volumes  
214 of avian plasma for neonicotinoids and the metabolite 6-CNA using an easy (approximately one  
215 hour) sample preparation approach combining quick deproteinization and dilution. Plasma  
216 samples were mixed with methanol solution(s) and acidified with formic acid, and blown down  
217 to dryness to denature biomolecule components in the plasma as much as possible, then  
218 neonicotinoids were re-dissolved in 500  $\mu$ L 20:80 methanol:water and separated from the

219 precipitates by centrifuge. It was found that an increase or decrease in temperature before blown  
220 down in an attempt to help with the denaturation process did not improve recovery.

221         Based on recoveries of fortified isotope-labelled neonicotinoids in commercial bovine  
222 and chicken plasma with different dilution factors (2.5 to 10), we found that dilution helped  
223 reduce matrix effects. At a dilution factor of 2.5 (200  $\mu$ L plasma in 500 final volume), signals for  
224 all four isotope-labelled neonicotinoids were extremely suppressed for bovine plasma samples as  
225 shown in Figure 1. This signal suppression was also observed for acetamiprid-d<sub>3</sub>, clothianidin-  
226 d<sub>3</sub>, and thiamethoxam-d<sub>3</sub> from chicken plasma with the same dilution factor, but the level of  
227 suppression was much less. For imidacloprid-d<sub>4</sub>, its signal was actually enhanced in the 200  $\mu$ L  
228 chicken plasma sample. These results indicate that method performance can vary for plasma  
229 samples from different animal species. Dilution of bovine plasma helped to reduce matrix effects  
230 and the suppression caused by matrix components was improved significantly at the dilution  
231 factor of 10 (50  $\mu$ L plasma in 500  $\mu$ L final volume), which we used for all further analysis.

232         It is worth pointing out that imidacloprid was detected at ~13 pg/mL in the commercial  
233 chicken plasma sample during the analysis, which indicated that there was exposure to this  
234 commonly used insecticide.

235

### 236 **3.2 Songbird Plasma Method Development and Validation**

237         Following method optimization, captive white-crowned sparrow plasma pool samples  
238 were confirmed to contain no detectable neonicotinoids, and were then used for final validations  
239 (Table 2). The average recoveries for target compounds in spiked sparrow plasma ranged from  
240 95.7 to 101.3% and the recoveries for the four isotope-labelled neonicotinoids were in the range  
241 of 58.0 to 82.6%. Clothianidin-d<sub>3</sub> suffered the worst ion suppression among the four isotope-

242 labelled neonicotinoids in fortified samples while acetamiprid-d<sub>3</sub> was affected the least. It is still  
243 not clear why certain target compounds such as clothianidin suffer more from matrix effects, but  
244 this type of ion suppression observed during LC-MS/MS analysis of neonicotinoids in biological  
245 samples could be minimized by solid phase extraction, cleanup and ultra performance liquid  
246 chromatography (UPLC) separation [55]. Internal standard correction was used in this method to  
247 compensate for sample preparation loss, matrix effects and instrument variability. The method  
248 detection limits are in the low pg/mL range for all parent neonicotinoids and ~178 pg/mL for 6-  
249 CNA. Only 50 µL of blood plasma was needed for the analysis and the sample preparation work  
250 of deproteinization and dilution was minimal (approx. 1 hour). The isotope-labelled internal  
251 standards (ILISs) acetamiprid-d<sub>3</sub>, clothianidin-d<sub>3</sub>, imidacloprid-d<sub>4</sub> and thiamethoxam-d<sub>3</sub> were  
252 used to correct extraction loss, matrix effects and instrument variability and to monitor method  
253 performance in each sample. Relative standard deviation was around 5% for the four labelled  
254 internal standards, and less than 2.5% for all target compounds after the internal standard  
255 correction (Table 2). As mentioned earlier, there was only one study published before on the  
256 measurement of neonicotinoids in bird blood samples [42]. The performance data demonstrate  
257 that the recoveries of target compounds and the relative standard deviation of the recoveries are  
258 improved compared to the previous study, and more importantly, the method would provide  
259 precise and accurate results at low part-per-trillion (pg/mL) levels for neonicotinoid pesticides  
260 monitored in bird plasma.

### 261 **3.3 Evaluation of Neonicotinoid Exposure in Free Living Songbirds**

262 The analytical procedure developed was applied to analyze plasma samples from free-  
263 living white-crowned sparrows captured in southern Ontario, Canada during spring migration in  
264 May 2017, to assess their exposure to neonicotinoid insecticides. We detected four

265 neonicotinoids, acetamiprid, imidacloprid, thiacloprid and thiamethoxam (Table 3). One sample  
266 (sample ID # 148) contained all four neonicotinoids. Figure 2 shows the chromatograms of the  
267 quantitation ion of each analyte in this sample, alongside the ion chromatograms in middle level  
268 calibration standard for comparison. Imidacloprid was above the MDL in 28 of the 36 plasma  
269 samples (78%), with the highest concentration of 177 pg/mL. The other eight samples also  
270 contained trace amount of imidacloprid but their levels were less than the MDL of 4.6 pg/mL.  
271 Eight of the 36 samples (22%) showed positive results for thiamethoxam with the highest  
272 concentration of 33.7. Trace amount of thiamethoxam was detected in another three samples at  
273 less than MDL (4.5 pg/mL) levels. Four samples showed positive results for acetamiprid (11%)  
274 and thiacloprid (11%), respectively, but all were below 5 pg/mL. Six samples showed trace  
275 amount of clothianidin (as demonstrated in the ion chromatogram in Figure 2A), a commonly  
276 used neonicotinoid in North America; however, there were none higher than the MDL level of  
277 7.4 pg/mL. We caution that this might be attributed to clothianidin having the worst ion  
278 suppression encountered during LC-MS/MS detection. No dinotefuran, flonicamid, nitenpyram  
279 or 6-CNA was detected in the blood samples of the wild-caught white-crowned sparrows.

280         Studies of neonicotinoid residues in avian blood are rare; the only other available study  
281 was done in Eurasian eagle owls (*Bubo bubo*), which are top predators and residues would most  
282 likely be through secondary exposure from consuming contaminated seed-eating prey (e.g.  
283 rabbits *Oryctolagus cuniculus*, partridges *Alectoris ruffa* and pigeons *Columba spp*). Thirty owls  
284 were tested for nine neonicotinoids, and only one owl had detectable residues for IMI only [42],  
285 although limits of detection were about three orders of magnitude above those in the present  
286 study.

287           The detection of circulating neonicotinoids in the majority of wild-caught sparrows in  
288 this study suggests that free-living seed-eating birds are being exposed to neonicotinoids at high  
289 enough concentrations and/or frequent enough exposures for the parent compounds to be  
290 routinely detected in circulation. The birds measured in this study were caught on a northerly  
291 migratory stopover between their wintering and breeding grounds. Eastern white-crowned  
292 sparrows overwinter from central Texas to the lower Ohio River Valley, and breed in boreal  
293 shrub habitat in north eastern Canada. Their diet is > 90% plant based, and includes a variety of  
294 seeds as well as small grains (e.g. oats, wheat, barley, corn) [56]. The timing of spring migration  
295 overlaps with spring seeding for many agricultural crops, and birds may be exposed through  
296 direct consumption of neonicotinoid treated seeds. It is also possible that exposure comes from  
297 consumption of contaminated water, soil or seeds produced by plants that have taken up  
298 neonicotinoids systemically. Neonicotinoids are taken up into plant tissues and can be  
299 translocated to the flowers and seeds [57]. Residues have been detected in crop plants, as well as  
300 in non-crop plants from field margins that are exposed to neonicotinoids through agricultural  
301 runoff or drift [58, 59]. Most plant residue analysis has focused on pollen and nectar with respect  
302 to bee exposure, while seed concentrations are rarely reported. In addition to seed coatings, there  
303 could be other usage such as spray applications of neonicotinoids to both agricultural and urban  
304 areas. Other possible routes of avian exposure include inhalation and dermal exposure to  
305 environmental residues or consuming neonicotinoid contaminated surface water. Further work is  
306 needed to determine exposure pathways in birds including field surveys of treated seed  
307 availability and consumption by wildlife, characterization of systemic seed concentrations, and  
308 comparison of circulating neonicotinoid concentrations in non-seed eating birds (e.g. aerial  
309 insectivores) that use agricultural areas.



310

### 311 **3.4 Controlled Imidacloprid Dosing Study**

312 Under controlled dosing via oral gavage, white-crowned sparrows were exposed to  
313 environmentally realistic concentrations of imidacloprid and concentration changes in blood  
314 levels were monitored. There was no significant difference in pre-dosing plasma concentrations  
315 among treatment groups ( $p \geq 0.815$ ). There was a dose-dependent increase in plasma  
316 imidacloprid, and a significant interaction between time and dose ( $F_{2,33} = 19.15, p < 0.0001$ ).  
317 Plasma imidacloprid concentrations significantly increased between pre- and post-dosing in the  
318 low ( $p < 0.0001$ ) and high dose ( $p < 0.0001$ ) birds, but not the control birds ( $p = 0.418$ ) (Fig. 3).  
319 There was a large variation in the plasma imidacloprid concentration at 6 hours post-dosing (low  
320 range = 100 to 3,210 pg/mL; high range = 840 to 35,100 pg/mL), suggesting individual variation  
321 in uptake, biotransformation and/or elimination; however, all birds dosed with imidacloprid had  
322 higher concentrations post-dosing than pre dosing, indicating that imidacloprid can be detected if  
323 a blood sample is taken within the 6-hour time frame after ingesting a similar mass of  
324 imidacloprid as found on a single treated seed. Interestingly, the geometric mean plasma  
325 concentration from the high dose was 24x higher than the low dose (5558.2 pg/mL vs 234.2  
326 pg/mL), despite the difference in administered dose being only 3.25x higher (3.9 mg/kg bw vs  
327 1.2 mg/kg bw). This pattern suggests that at concentrations in the high dose range, the rate of  
328 biotransformation and elimination is more limited. The timing and concentration of exposures  
329 are likely difficult to establish from blood residues; however, plasma concentrations in the range  
330 detected here (0.1 to >100 pg/mL) would likely indicate recent exposure. The majority of  
331 sparrows had pre-dosing concentrations below this level, but three birds had plasma

332 concentrations >100 pg/mL at capture. In comparison, the study of neonicotinoid residues in  
333 Eurasian eagle owl blood reported 3280 pg/mL of imidacloprid in one individual [42].

334 Improved knowledge of neonicotinoid toxicokinetics in birds, including a time-course  
335 analysis of plasma following controlled exposures combined with quantification of metabolites,  
336 could help further characterize whether residues detected from wild bird samples are due to  
337 recent low-level exposures or to reduced levels from past exposures. The metabolite 6-CNA was  
338 not detected in any samples pre- or post-dosing. Other metabolites not included in this study,  
339 such as 5-OH-imidacloprid, imidacloprid-olefin, and imidacloprid-nitrosoimine could be useful  
340 to quantify in the future to help understand the toxicokinetics of neonicotinoids in songbirds.

#### 341 **4. Conclusion**

342 An efficient deproteinization and “dilute and shoot” LC-MS/MS approach for the  
343 determination of eight neonicotinoid insecticides and metabolite 6-CNA in bird plasma are  
344 described in this study. Two MRM transitions with each target compound were monitored for  
345 their unambiguous identification in complex biological samples and four isotope-labelled  
346 internal standards were employed to improve data quality. The ease of sample preparation, the  
347 small 50 µL volume of plasma required and the low MDLs in the ranges of 2.3 to 15.9 pg/mL for  
348 the parent compounds make it a promising method for assessing exposure to neonicotinoid  
349 insecticides in songbirds or other small animals by non-lethal blood sampling. This sensitive  
350 method also provided a useful tool to characterize toxicologically and environmentally relevant  
351 exposures of neonicotinoids in avian wildlife even at low concentrations. Using this new method,  
352 this is the first study to confirm migratory songbirds have been exposed to four of the registered  
353 neonicotinoids: acetamiprid, imidacloprid, thiacloprid and thiamethoxam in North America. This

354 novel method will greatly complement future ecotoxicological research on avian wildlife  
355 exposure and effects to neonicotinoid insecticides, given their prevalence in agriculture  
356 worldwide.

357

### 358 **Acknowledgements**

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361 Kerry Peru for dosing solution analysis at National Hydrology Research Centre, Environment  
362 and Climate Change Canada was greatly appreciated. We also thank Amy Wilson for field  
363 assistance, and Bird Studies Canada and Long Point Bird Observatory for field site access and  
364 logistical support.

365

366 **Table 1.** Multiple reaction monitoring (MRM) transitions, corresponding collision energies  
 367 (CEs), isotope-labelled internal standards (ILISs) used, and calibration range for each target  
 368 compound (a, b denote two MRM transitions monitored for a given target compound, MRM a  
 369 was used for quantitation and MRM b was used for confirmation).

MRM Transition ID	Q1 Mass	Q3 Mass	CE (eV)	ILIS	Calibration Range (ng/L)
6-CNA a	158	122	25	imidacloprid-d <sub>4</sub>	40-4000
6-CNA b	158	78	33	imidacloprid-d <sub>4</sub>	40-4000
acetamiprid a	223	126	29	acetamiprid-d <sub>3</sub>	2-200
acetamiprid b	223	90	46	acetamiprid-d <sub>3</sub>	2-200
clothianidin a	250	169	18	clothianidin-d <sub>3</sub>	2-200
clothianidin b	250	132	19	clothianidin-d <sub>3</sub>	2-200
dinotefuran a	203	129	15	clothianidin-d <sub>3</sub>	2-200
dinotefuran b	203	114	15	clothianidin-d <sub>3</sub>	2-200
flonicamid a	230	203	25	clothianidin-d <sub>3</sub>	4-400
flonicamid b	230	174	25	clothianidin-d <sub>3</sub>	4-400
imidacloprid a	256	209	23	imidacloprid-d <sub>4</sub>	2-200
imidacloprid b	256	175	25	imidacloprid-d <sub>4</sub>	2-200
nitenpyram a	271	126	38	clothianidin-d <sub>3</sub>	2-200
nitenpyram b	271	224	20	clothianidin-d <sub>3</sub>	2-200
thiacloprid a	253	126	29	acetamiprid-d <sub>3</sub>	1-100
thiacloprid b	253	90	50	acetamiprid-d <sub>3</sub>	1-100
thiamethoxam a	292	211	19	thiamethoxam-d <sub>3</sub>	2-200
thiamethoxam b	292	181	33	thiamethoxam-d <sub>3</sub>	2-200
acetamiprid-d <sub>3</sub>	226	126	29		
clothianidin-d <sub>3</sub>	253	172	19		
imidacloprid-d <sub>4</sub>	260	213	23		
thiamethoxam-d <sub>3</sub>	295	214	19		

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372 **Table 2.** Spiking levels in 50  $\mu$ L fortified plasma pool samples (n = 12) from captive white-  
 373 crowned sparrows, final concentrations in reconstituted solutions for instrument analysis,  
 374 average recovery (Avg Rec%), standard deviation (SD), relative standard deviation (RSD%) and  
 375 calculated method detection limits (MDLs) for target analytes.  
 376

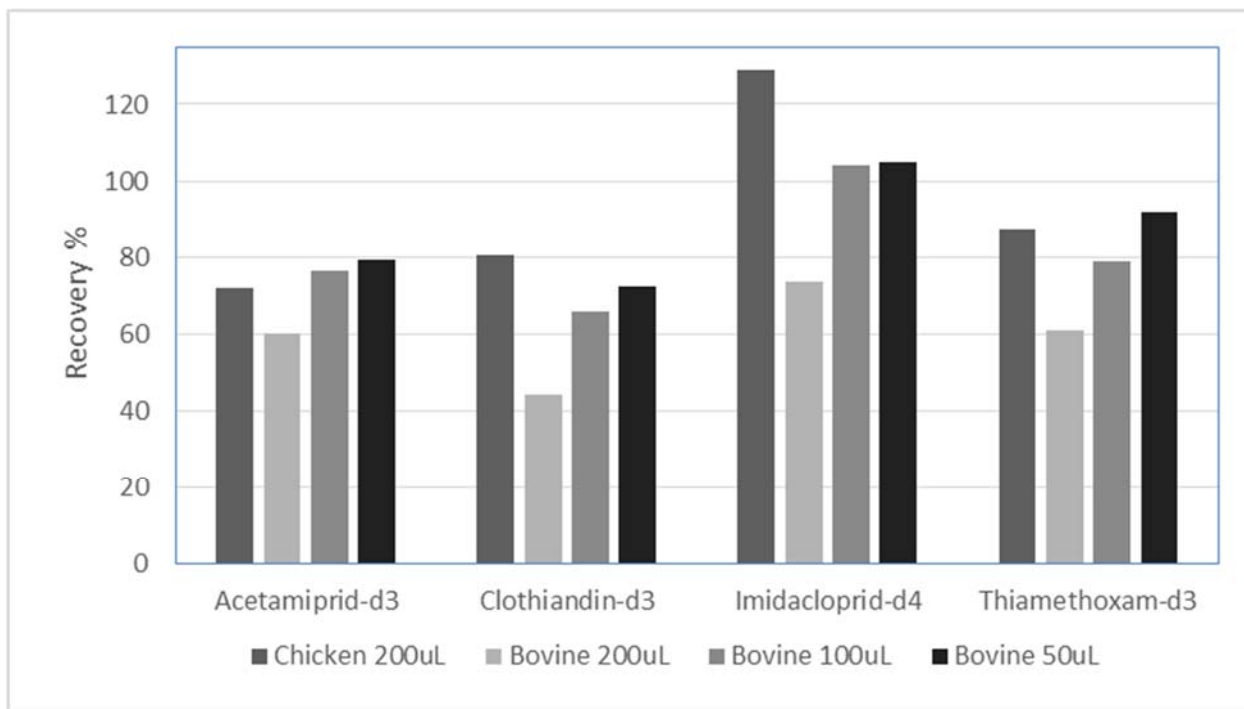
Compound Name	Spiking Level ng/L	Final Concentration ng/L	Avg Rec %	RSD %	MDL ng/L
6-CNA	3200	320	101.3	2.0	177.7
acetamiprid	160	16	96.9	0.8	3.6
clothianidin	160	16	98.5	1.7	7.4
dinotefuran	160	16	95.7	2.1	9.1
flonicamid	320	32	99.4	1.8	15.9
imidacloprid	160	16	99.4	1.0	4.6
nitenpyram	160	16	101.2	1.9	8.8
thiacloprid	80	8	97.0	1.1	2.3
thiamethoxam	160	16	98.9	1.0	4.5
acetamiprid-d <sub>3</sub>	400	40	82.6	4.5	
clothianidin-d <sub>3</sub>	400	40	58.0	4.3	
imidacloprid-d <sub>4</sub>	400	40	61.9	5.3	
thiamethoxam-d <sub>3</sub>	400	40	70.2	4.7	

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380 **Table 3.** Calculated analyte concentrations (pg/ml) of acetamiprid, imidacloprid, thiacloprid and  
 381 thiamethoxam in 36 plasma samples of white-crowned sparrows at capture. Clothianidin,  
 382 dinotefuran, flonicamid, nitenpyram and 6-CNA were non-detected (ND) in all samples.  
 383

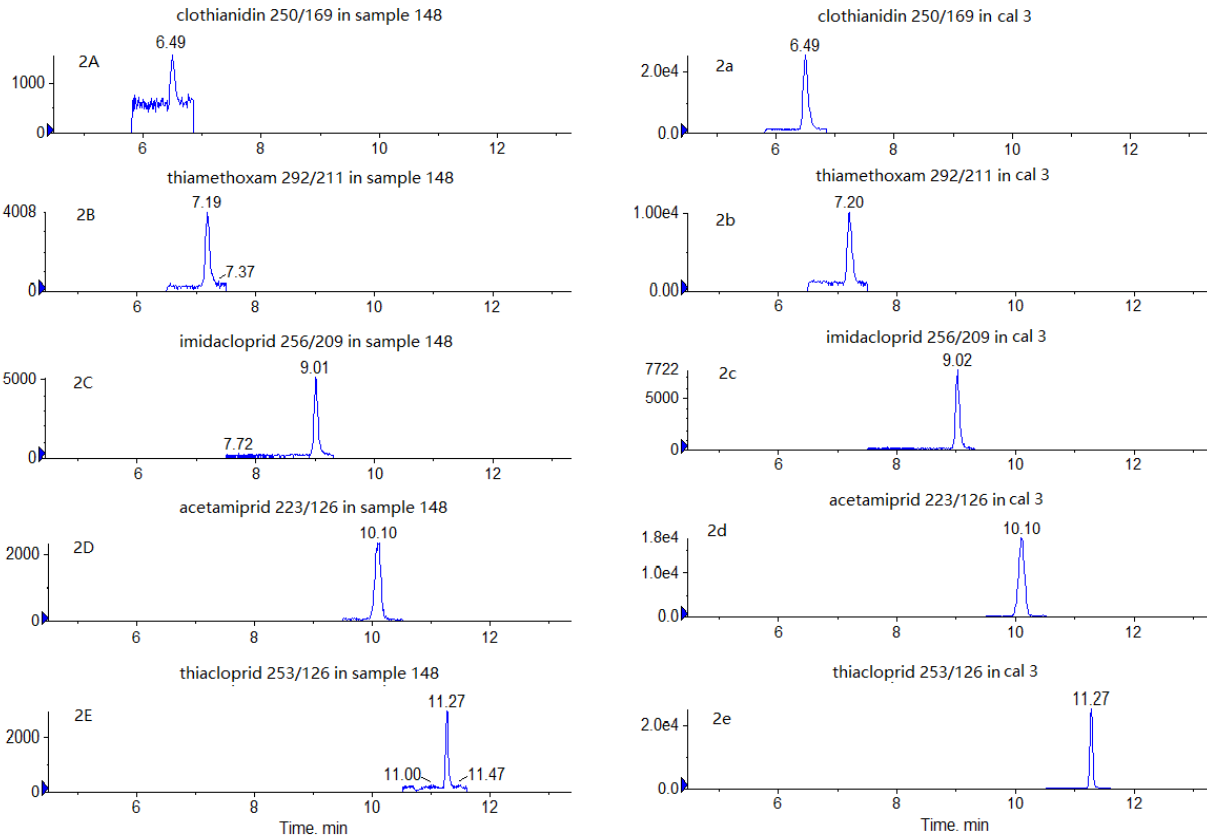
Sample ID	Sampling Date	Acetamiprid	Imidacloprid	Thiacloprid	Thiamethoxam
189	8-May	ND	7.2	ND	21.6
869	8-May	ND	<MDL	ND	7.2
937	8-May	ND	9.9	ND	10.7
8154	8-May	ND	5.0	ND	33.7
195	9-May	ND	5.1	ND	10.4
204	9-May	ND	8.7	ND	ND
216	9-May	3.7	23.1	2.6	ND
220	9-May	ND	9.5	ND	7.4
842	9-May	ND	7.4	ND	<MDL
858	9-May	ND	<MDL	ND	ND
888	9-May	ND	36.1	ND	ND
889	9-May	ND	<MDL	ND	ND
943	9-May	ND	6.4	ND	ND
944	9-May	ND	16.5	ND	ND
949	9-May	ND	8.1	ND	ND
950	9-May	ND	<MDL	ND	ND
955	9-May	ND	<MDL	ND	ND
148	10-May	3.9	34.7	2.6	11.6
151	10-May	ND	129.0	ND	ND
114	11-May	ND	105.0	<MDL	<MDL
150	11-May	ND	177.0	ND	ND
155	11-May	3.9	83.0	2.5	ND
117	12-May	3.8	55.8	3.1	ND
575	12-May	ND	58.7	ND	5.1
180	13-May	ND	20.3	<MDL	ND
553	13-May	ND	61.2	ND	ND
560	13-May	ND	12.8	ND	ND
569	13-May	ND	7.2	ND	ND
573	13-May	ND	<MDL	ND	ND
152	14-May	ND	22.1	ND	ND
283	14-May	ND	63.9	ND	ND
411	14-May	ND	<MDL	ND	ND
551	14-May	ND	15.6	<MDL	ND
571	14-May	ND	5.7	ND	ND
576	14-May	ND	<MDL	ND	<MDL
3154	14-May	ND	8.1	ND	ND
# Samples ≥ MDLs		4	28	4	8
Max conc.		3.9	177	3.1	33.7

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**Figure 1.** Recoveries of four isotope-labelled neonicotinoids in fortified samples with dilution factor of 2.5 (200 µL) for chicken plasma, dilution factors of 2.5 (200 µL), 5 (100 µL), and 10 (50 µL) for bovine plasma.



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393

394 **Figure 2.** Ion chromatograms of clothianidin 250/169 in sample 148 (A left ) and middle level

395 calibration standard solution cal 3 (a right), thiamethoxam 292/211 in sample 148 (B left) and

396 middle level calibration standard solution cal 3 (b right), imidacloprid 256/209 in sample 148 (C

397 left) and middle level calibration standard solution cal 3 (c right), acetamiprid 223/126 in

398 sample 148 (D left) and middle level calibration standard solution cal 3 (d right), and thiacloprid

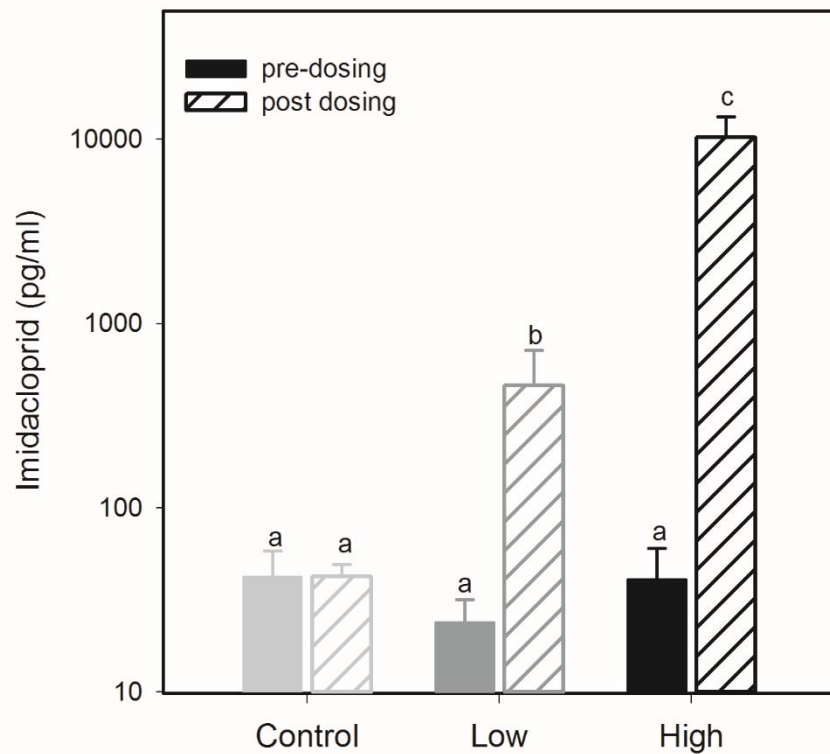
399 253/126 in sample 148 (E left) and middle level calibration standard solution cal 3 (e right).

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404 **Figure 3.** Mean ( $\pm$ SE) measured plasma imidacloprid concentrations ( $\log_{10}$  scale) at capture (pre-  
 405 dosing) and 6 hours after oral exposure (post-dosing) to either the vehicle control (sunflower oil),  
 406 1.2 (low) or 3.9 (high) mg imidacloprid per kg body weight. Significant differences ( $\alpha < 0.05$ )  
 407 between groups indicated with different lower case letters.

- 
- [1] N. Simon-Delso, V. Amaral-Rogers, L. P. Belzunces, J. M. Bonmatin, M. Chagnon, C. Downs, L. Furlan, D. W. Gibbons, C. Giorio, V. Girolami, D. Goulson, D. P. Kreutzweiser, C. H. Krupke, M. Liess, E. Long, M. McField, P. Mineau, E. A. D. Mitchell, C. A. Morrissey, D. A. Noome, L. Pisa, J. Settele, J. D. Stark, A. Tapparo, H. Van Dyck, J. Van Praagh, J. P. Van der Sluijs, P. R. Whitehorn, M. Wiemers, Systemic insecticides (neonicotinoids and fipronil): trends, uses, mode of action and metabolites. *Environ. Sci. Pollut. Res.* 2015, 22: 5–34.
- [2] R. Nauen, U. Ebbinghaus-Kintscher, R. Schmuck, Toxicity and nicotinic acetylcholine receptor interaction of imidacloprid and its metabolites in *Apis mellifera* (Hymenoptera: Apidae). *Pest Manag. Sci.* **2001**, 57: 577-586.
- [3] M. Tomizawa, J.E. Casida, Neonicotinoid insecticide toxicology: mechanisms of selective action. *Annu. Rev. Pharmacol. Toxicol.* **2005**, 45: 247-268.
- [4] M.J. Palmer, C. Moffat, N. Saranzewa, J. Harvey, G. A. Wright, C. N. Connolly, Cholinergic pesticides cause mushroom body neuronal inactivation in honeybees. *Nat. Commun.* **2013**, 4: 1634. <http://dx.doi.org/10.1038/ncomms2648>.
- [5] M. C. Cavallaro, C. A. Morrissey, J. V. Headley, K. M. Peru, K Liber, Comparative chronic toxicity of imidacloprid, clothianidin, and thiamethoxam to *Chironomus dilutus* and estimation of toxic equivalency factors. *Environ. Toxicol. Chem.* **2017**, 36: 372–382. doi:10.1002/etc.3536
- [6] P. J. Van den Brink, J. M. Van Smeden, R. S. Bekele, W. Dierick, D. M. De Gelder, M. Noteboom, I. Roessink. Acute and chronic toxicity of neonicotinoids to nymphs of a mayfly species and some notes on seasonal differences. *Environ. Toxicol. Chem.* **2016**, 35: 128–133. doi:10.1002/etc.3152

- 
- [7] E. M. Maloney, C. A. Morrissey, J. V. Headley, K. M. Peru, K. Liber, Cumulative toxicity of neonicotinoid insecticide mixtures to *Chironomus dilutus* under acute exposure scenarios. *Environ. Toxicol. Chem.* **2017**, 36: 3091–3101.
- [8] M. Gross, Pesticides linked to bee deaths. *Curr. Biol.* **2008**, 18: R684.
- [9] A. Decourtye, J. Devillers, Ecotoxicity of neonicotinoid insecticides to bees. *Adv. Exp. Med. Biol.* **2010**, 683: 85-95.
- [10] T. Blacquièrre, G. Smagghe, C. A. M. van Gestel, V. Mommaerts. Neonicotinoids in bees: a review on concentrations, side-effects and risk assessment. *Ecotoxicology* **2012**, 21: 973–992.
- [11] C. A. Hallmann, R. P. B. Foppen, C. A. M. van Turnhout, H. de Kroon & E. Jongejans, *Nature*, **2014**, 511: 341-343
- [12] M. L. Eng, B. J. M. Stutchbury, C. A. Morrissey, Imidacloprid and chlorpyrifos insecticides impair migratory ability in a seed-eating songbird, *Sci. Rep.* **2017**, 7: 15176
- [13] F. Millot, A. Decors, O. Mastain, T. Quintaine, P. Berny, D. Vey, R. Lasseur, E. Bro, Field evidence of bird poisonings by imidacloprid-treated seeds: a review of incidents reported by the French SAGIR network from 1995 to 2014, *Environ. Sci. Pollut. Res.* **2017**, 24: 5469–5485
- [14] A. Lopez-Antia, J. Feliu, P. R. Camarero, M. E. Ortiz-Santaliestra, R. Mateo, Risk assessment of pesticide seed treatment for farmland birds using refined field data, *J. Appl. Ecol.* **2016**, 53: 1373–1381
- [15] H. M. H. Ertl, M. A. Mora, D. J. Brightsmith, J. A. Navarro-Alberto, Potential impact of neonicotinoid use on Northern bobwhite (*Colinus virginianus*) in Texas: A historical analysis, *PLoS ONE* **2018**, 13: e0191100

- 
- [16] Government of Ontario. Neonicotinoid regulations, 30 January 2017. Retrieved 03 April 2018, from <https://www.ontario.ca/page/neonicotinoid-regulations>
- [17] EFSA (European Food Safety Authority), 2018. Evaluation of the data on clothianidin, imidacloprid and thiamethoxam for the updated risk assessment to bees for seed treatments and granules in the EU. EFSA supporting publication 2018:EN-1378. 31 pp. doi:10.2903/sp.efsa.2018.EN-1378
- [18] J.C. Anderson, C. Dubetz, V.P. Palace, Neonicotinoids in the Canadian aquatic environment: A literature review on current use products with a focus on fate, exposure, and biological effects. *Sci. Total Environ.* **2015**, 505: 409-422.
- [19] W. J. Sacks, D. Deryng, J. A. Foley, N. Ramankutty, [Crop planting dates: an analysis of global patterns](#), *Global Ecol, Biogeogr.* **2010**, 19: 607-620
- [20] U. Turaga, S. T. Peper, N. R. Dunham, N. Kumar, W. Kistler, S. Almas, S. M. Presley, R. J. Kendall, A survey of neonicotinoid use and potential exposure to northern bobwhite (*Colinus virginianus*) and scaled quail (*Callipepla squamata*) in the Rolling Plains of Texas and Oklahoma, *Environ. Toxicol. Chem.* **2016**, 35: 511–1515
- [21] A. M. MacDonald, C. M. Jardine, P. J. Thomas, N. M. Nemeth, Neonicotinoid detection in wild turkeys (*Meleagris gallopavo silvestris*) in Ontario, Canada. *Environ. Sci. Pollut. Res.*, **2018**, 25:1-7
- [22] S. Espín, A. J. J. García-Fernández, D. Herzke, R.F.F. Shore, B. van Hattum, E. Martínez-López, M. Coeurdassier, I. Eulaers, C. Fritsch, P. Gómez-Ramírez, V. L. B. L. B. Jaspers, O. Krone, G. Duke, B. Helander, R. Mateo, P. Movalli, C. Sonne, N. W. W. van den Brink, Tracking pan-continental trends in environmental contamination using sentinel raptors-what

---

types of samples should we use? *Ecotoxicology*, **2016**, 25: 777-801. doi:10.1007/s10646-016-1636-8

[23] J. C. Owen, Collecting, processing, and storing avian blood: a review, *J. Field Ornithol.* **2011**, 82: 339–354

[24] M. M. Galera, A. G. Frenich, J. L. M. Vidal, P. P. Vázquez, Resolution of imidacloprid pesticide and its metabolite 6-chloronicotinic acid using cross-sections of spectrochromatograms obtained by high-performance liquid chromatography with diode-array detection. *J. Chromatogr. A* **1998**, 799: 149-154.

[25] S. Seccia, P. Fidente, D. A. Barbini, P. Morrica, Multiresidue determination of nicotinoid insecticide residues in drinking water by liquid chromatography with electrospray ionization mass spectrometry. *Anal. Chim. Acta* **2005**, 553: 21-26.

[26] A. M. Rodrigues, V. Ferreira, V. V. Cardoso, E. Ferreira, M. J. Benoliel, Determination of several pesticides in water by solid-phase extraction, liquid chromatography and electrospray tandem mass spectrometry. *J. Chromatogr. A* **2007**, 1150: 267-278.

[27] K. Starner, K. S. Goh. Detections of the neonicotinoid insecticide imidacloprid in surface waters of three agricultural regions of California, USA, 2010–2011. *Bull. Environ. Contam. Toxicol.* **2012**, 88: 316-321.

[28] M. L. Hladik, D. W. Kolpin, K. M. Kuivila, Widespread occurrence of neonicotinoid insecticides in streams in a high corn and soybean producing region, USA. *Environ. Poll.* **2014**, 193: 189-196.

- 
- [29] F. Sánchez-Bayo, R. V. Hyne, Detection and analysis of neonicotinoids in river waters – Development of a passive sampler for three commonly used insecticides. *Chemosphere* **2014**, 99, 143-151.
- [30] A. R. Main, J. V. Headley, K. M. Peru, N. L. Michel, A. J. Cessna, C. A. Morrissey, Widespread use and frequent detection of neonicotinoid insecticides in wetlands of Canada's prairie pothole region. *PLoS ONE* **2014**, 9: e92821.
- [31] A. Jones, P. Harrington, G. Turnbull, Neonicotinoid concentrations in arable soils after seed treatment applications in preceding years. *Pest. Manag. Sci.* **2014**, 70: 1780-1784.
- [32] E. Dankyi, C. Gordon, D. Carboo, I. S. Fomsgaard, Quantification of neonicotinoid insecticide residues in soils from cocoa plantations using a QuEChERS extraction procedure and LC-MS/MS. *Sci. Total Environ.* **2014**, 499: 276-283.
- [33] E. Dankyi, D. Carboo, C. Gordon, I. S. Fomsgaard, Application of the QuEChERS procedure and LC-MS/MS for the assessment of neonicotinoid insecticide residues in cocoa beans and shells. *J. Food Compos. Anal.* **2015**, 44: 149-157.
- [34] E. Dankyi, C. Gordon, D. Carboo, V. A. Apalangya, I. S. Fomsgaard, Sorption and degradation of neonicotinoid insecticides in tropical soils. *J. Environ. Sci. Health B (Pesticides, food contaminants, and agricultural wastes)* **2018**, 1-8. DOI: 10.1080/03601234.2018.1473965
- [35] A. Kamel, Refined methodology for the determination of neonicotinoid pesticides and their metabolites in honey bees and bee products by liquid chromatography–tandem mass spectrometry (LC-MS/MS). *J. Agric. Food Chem.* **2010**, 58: 5926-5931

- 
- [36] W. Xie, C. Han, Y. Qian, H. Ding, X. Chen, J. Xi, Determination of neonicotinoid pesticides residues in agricultural samples by solid-phase extraction combined with liquid chromatography–tandem mass spectrometry. *J. Chromatogr. A* **2011**, 1218: 4426-4433.
- [37] S. Liu, Z. Zheng, F. Wei, Y. Ren, W. Gui, H. Wu, G. Zhu, Simultaneous determination of seven neonicotinoid pesticide residues in food by ultraperformance liquid chromatography tandem mass spectrometry. *J. Agric. Food Chem.* **2010**, 58: 3271-3278.
- [38] Z. Xiao, X. Li, X. Wang, J. Shen, S. Ding, Determination of neonicotinoid insecticides residues in bovine tissues by pressurized solvent extraction and liquid chromatography–tandem mass spectrometry. *J. Chromatogr. B* **2011**, 879: 117-122.
- [39] A. Kamel, Y. Qian, E. Kolbe, C. Stafford, Development and validation of a multiresidue method for the determination of neonicotinoid and macrocyclic lactone pesticide residues in milk, fruits, and vegetables by ultra-performance liquid chromatography/MS/MS. *J. AOAC Inter.* **2010**, 93: 389-399.
- [40] M. Chen, E. M. Collins, L. Tao, C. Lu, Simultaneous determination of residues in pollen and high-fructose corn syrup from eight neonicotinoid insecticides by liquid chromatography–tandem mass spectrometry. *Anal. Bioanal. Chem.* **2013**, 405: 9251-9264.
- [41] P. Jovanov, V. Guzsvány, M. Franko, Sa. Lazić, M. Sakač, B. Šarić, V. Banjac, Multi-residue method for determination of selected neonicotinoid insecticides in honey using optimized dispersive liquid-liquid microextraction combined with liquid chromatography-tandem mass spectrometry, *Talanta*, **2013**, 111: 125-133
- [42] A. Taliansky-Chamudis, P. Gómez-Ramírez, M. León-Ortega, A.J. García-Fernández. Validation of a QuEChERS method for analysis of neonicotinoids in small volumes of blood and

---

assessment of exposure in Eurasian eagle owl (*Bubo bubo*) nestlings. *Sci. Total Environ.* **2017**, 595: 93–100

[43] U. Kapoor, M. K. Srivastava, P. Trivedi, V. Garg, L. P. Srivastava, Disposition and acute toxicity of imidacloprid in female rats after single exposure. *Food Chem. Toxicol.* **2014**, 68:190–195.

[44] T. C. Marrs, 2012. Mammalian toxicology of insecticides. Issues in Toxicology. Roy. Soc. Chem., Cambridge

[45] C. Hao, M. R. Noestheden, X. Zhao, D. Morse, Liquid chromatography/tandem mass spectrometry analysis of neonicotinoid pesticides and 6-chloronicotinic acid in environmental water with direct aqueous injection. *Anal. Chim. Acta* **2016**, 925: 43-50

[46] G. Tanner, C. Czerwenka, LC-MS/MS Analysis of Neonicotinoid Insecticides in Honey: Methodology and Residue Findings in Austrian Honeys, *J. Agric. Food Chem.* **2011**, 59: 12271-12277.

[47] S. Lazić, D. Šunjka, N. Grahovac, V. Guzsvány, F. Bagi, D. Budakov, Application of Liquid Chromatography with Diode-Array Detector for Determination of Acetamiprid and 6-chloronicotinic Acid Residues in Sweet Cherry Samples, *Pestic. Phytomed. (Belgrade)* **2012**, 27: 321–329.

[48] S. Totti, M. Fernández, S. Ghini, Y. Picó, F. Fini, J. Mañes, S. Girotti, Application of matrix solid phase dispersion to the determination of imidacloprid, carbaryl, aldicarb, and their main metabolites in honeybees by liquid chromatography–mass spectrometry detection, *TALANTA* **2006**, 69: 724-729.



- 
- [49] H. John, M. Eddleston, R. E. Clutton, F. Worek, H. Thiermann, Simultaneous quantification of the organophosphorus pesticides dimethoate and omethoate in porcine plasma and urine by LC-ESI-MS/MS and flow-injection-ESI-MS/MS. *J. Chromatogr. B* **2010**, 878:1234-1245
- [50] H. Bagheri, A. Es'haghi, A. Es-haghi, N. Mesbahi, A high-throughput approach for the determination of pesticide residues in cucumber samples using solid-phase microextraction on 96-well plate, [Anal. Chim. Acta](#) **2012**, 740: 36-42. doi: 10.1016/j.aca.2012.06.001
- [51] F. J. [Schenck](#) J. E. [Hobbs](#), "Evaluation of the quick, easy, cheap, effective, rugged, and safe (QuEChERS) approach to pesticide residue analysis," [Bull. Environ. Contam. Toxicol.](#) **2004**, 73: 24-30.
- [52] S. Esposito, E. Bracacel, M. Nibbio, R. Speziale, L. Orsatti, M. Veneziano, E. Monteagudo, F. Bonelli, "Use of 'dilute-and-shoot' liquid chromatography-high resolution mass spectrometry in preclinical research: Application to a DMPK study of perhexiline in mouse plasma," *J. Pharm. Biomed. Anal.* **2016**, 118: 70–80.
- [53] T. R. Stafford, NTN 33893 1.5G: an acute oral LD 50 with house sparrows, *Passer domesticus*. *Mobay Corporation, Kansas City, Missouri Report No. 101324. 23pp* (1991)
- [54] Federal Register, U.S. Code of Federal Regulations, Part 136, Appendix B, 49 FR 43430, Oct. 26, 1984; 50 FR 694, 696, Jan. 4, 1985, as amended at 51 FR 23703, June 30, 1986
- [55] Q. [Zhang](#), X. [Wang](#), Z. [Li](#), H. [Jin](#), Z. [Lu](#), C. [Yu](#), Y. F. [Huang](#), M. [ZhaoWang](#), Simultaneous determination of nine neonicotinoids in human urine using isotope-dilution ultra-performance liquid chromatography-tandem mass spectrometry, *Environ. Pollut.* **2018**, 240: 647-652

- 
- [56] G. Chilton, M. C. Baker, C. D. Barrentine, M. A. Cunningham, White-crowned Sparrow (*Zonotrichia leucophrys*), version 2.0. In *The Birds of North America* (P. G. Rodewald, editor). 1995, Cornell Lab of Ornithology, Ithaca, New York, USA. <https://doi.org/10.2173/bna.183>
- [57] F. M. Laurent, E. Rathahao, Distribution of [(14)C]imidacloprid in sunflowers (*Helianthus annuus* L.) following seed treatment, *J Agric Food Chem.* **2003**, 51: 8005-8010.
- [58] J. M. Bonmatin, C. Giorio, V. Girolami, D. Goulson, D. P. Kreutzweiser, C. Krupke, M. Liess, E. Long, M. Marzaro, E. A. D. Mitchell, D. A. Noome, N. Simon-Delso, A. Tapparo, Environmental fate and exposure; neonicotinoids and fipronil, *Environ. Sci. Pollut. Res.*, **2015**, 22: 35–67
- [59] C. Botías, A. David, E. M. Hill, D. Goulson, Contamination of wild plants near neonicotinoid seed-treated crops, and implications for non-target insects, *Sci. Total Environ.* **2016**, 566-567: 269-278

**Table S1.** Comparison of Imidacloprid low (1.2 mg IMI/kg bw/day) and high (3.9 mg IMI/kg bw/day) doses to potential exposure through seed consumption based on current application rates in Canada<sup>a</sup> and the US<sup>b</sup>. Estimates are based on an average white-crowned sparrow body mass of 27 g, with average seed consumption 7.9 g/day<sup>c</sup>

	Cereals <sup>e</sup> US	Cereals Canada	Canola US	Canola Canada	Corn US	Corn Canada	Sunflower US
Average mass of individual seed (g) <sup>d</sup>	0.035	0.035	0.003	0.003	0.38	0.38	0.05
Application rate	up to 2.4 fl oz/ 100lbs	up to 30 g a.i./ 100kg seed	up to 25.6 fl oz/ 100lbs	802 g a.i./ 100kg seed	up to 1.34 mg/ kernal	48 g a.i./ 80,000 seeds	up to 0.5 mg a.i./ seed
mg active ingredient (a.i.)/g seed	0.93	0.3	9.97	8.02	3.53	1.58	10
mg a.i. per individual seed	0.033	0.011	0.03	0.024	1.34	0.6	0.5
Number of seeds consumed per day (if 100% of diet)	226	226	2633	2633	21	21	158
mg IMI per day if diet 100% treated seeds	7.5	2.5	79	63.2	28.1	12.6	79
Number of seeds that contain equivalent to high dose (0.105 mg for a 27 g bird)	3.2	9.6	3.5	4.4	0.1	0.2	0.2
% of diet made up of treated seeds to consume equivalent to high dose	1.42	4.25	0.13	0.17	0.48	0.95	0.13
Number of seeds that contain equivalent to low dose(0.032 mg for a 27 g bird)	1	2.9	1.1	1.4	0.02	0.1	0.1
% of diet made up of treated seeds to consume equivalent to low dose	0.44	1.28	0.04	0.05	0.1	0.48	0.06

<sup>a</sup>Health Canada, Imidacloprid: Proposed Re-evaluation Decision. Pest Management Regulatory Agency PRVD2016-20 appendix IIa.

<sup>b</sup>Gaucho 600 flowable, US EPA reg 264-968

<sup>c</sup>food consumption values from M. Ramenofsky, R. Agatsuma, T. Ramfar, Environmental conditions affect the behavior of captive, migratory white-crowned sparrows. *The Condor*, 2008, 110:658-671.

<sup>d</sup>Seed masses from P. Mineau, C.Palmer, 2013. The impact of the nation's most widely used insecticides on birds. American Bird Conservancy

<sup>e</sup>wheat, barley, oats, rye, triticale